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1.0 Introduction

1.1 PURPOSE

The goal of this TMDL is to quantify the pollutant reductions needed to meet the water quality standards for chloride in Shingle Creek. The Shingle Creek TMDL for chloride is being established in accordance with Section 303(d) of the Clean Water Act, because the State of Minnesota has determined waters in the Shingle Creek Watershed exceed the State established standards for chloride.

1.2 PROBLEM IDENTIFICATION

Shingle Creek has an urban/suburban watershed located in the northwestern portion of the Minneapolis metropolitan region. The Creek is heavily used for stormwater management. The drainage system is composed of Shingle Creek, which is the major waterway, several tributaries, some intermittent streams, and a few man-made ditches. The main stem of Shingle Creek begins in Brooklyn Park in northwestern Hennepin County and flows generally southeast to its confluence with the Mississippi River in Minneapolis. Shingle Creek is formed at the junction of Bass Creek and Eagle Creek, two of the minor tributaries in the watershed. The creek is approximately 11 miles long and drops approximately 66 feet from its source to its mouth. Palmer Lake is the only lake directly on Shingle Creek.

High levels of chloride can directly harm aquatic organisms by disrupting natural osmo-regulatory processes. The MPCA has been actively developing plant and invertebrate indices of biological integrity (IBIs) in depressional wetlands to be used as indicators of wetland condition (Howard Markus, pers. comm.). As part of this research, standard water quality data are gathered in addition to biological data. Both the plant and invertebrate IBIs have been found to be
negatively correlated with chloride concentrations (Figure 1.1), suggesting that chloride may be causing declines in wetland diversity.

![Figure 1.1. Correlations of wetland plant (A) and invertebrate (B) IBIs with chloride concentration (* = P < 0.001).](image)

**Figure 1.1. Correlations of wetland plant (A) and invertebrate (B) IBIs with chloride concentration (\(*\) = P < 0.001).**

In 1998, Shingle Creek was listed on the Federal Clean Water Act’s 303(d) list of impaired waters for exceeding the chloride standard for aquatic life. The listing of Shingle Creek as impaired resulted from a limited sampling of chloride completed in 1996 by the US Geological Survey (USGS) at their discharge monitoring station at the Queen Avenue Bridge in Minneapolis. After reviewing the USGS data from Queen Avenue, the Shingle Creek Watershed Management Commission (SCWMC) has been sampling routinely for chloride in Shingle Creek. This TMDL was developed to address the 1998 listing for the impairment of aquatic life and recreation based on chloride exceedances.
Chloride is present in road salt, which most traffic authorities in the metropolitan area use extensively in the winter for snow and ice control. A network of freeways, highways, and local roads, all of which eventually drain to the creek, crisscross Shingle Creek’s watershed.

Section 303(d) of the Clean Water Act (CWA) requires the Minnesota Pollution Control Agency (MPCA) to identify waters that are not meeting State water quality standards and develop Total Maximum Daily Loads (TMDL) for those water bodies. A TMDL is the total amount of a pollutant that a water body can assimilate and still meet State water quality standards on a daily basis. Through the TMDL, pollutant loads can be distributed among the point and nonpoint sources in the watershed. These pollutant load allocations can then be used by managers to make science-based decisions on land use and management in the watershed.

In April 2002, the MPCA contracted with the Shingle Creek Watershed Management Commission, who subsequently contracted with Wenck Associates, Inc., to develop the TMDL for Chloride. The chloride TMDL included two phases: 1) field collection of data and 2) data analysis and TMDL modeling and allocation. The primary objectives pertinent to the Shingle Creek Chloride TMDL include:

- Define the spatial extent, persistence, and severity of chloride exceedances in the watershed,
- Identify and quantify the sources of chloride in Shingle Creek including point and nonpoint sources,
- Allocate Shingle Creek’s assimilative capacity to both point and nonpoint sources and develop safety margins protective of State water quality standards.

Since this TMDL represents the first TMDL for chloride in Minnesota, another aspect of this TMDL was the documentation of the lessons learned during this process. The concept for the lessons learned was to develop an understanding of chloride dynamics in a representative watershed to help provide key information region wide where it is likely that widespread
chloride exceedances may be occurring. The memo documenting lessons learned (Wenck 2004) was developed separately from this report.
2.0 Target Identification and Determination of Endpoints

2.1 IMPAIRED REACHES

In 1998, Shingle Creek was listed on the Federal Clean Water Act’s 303(d) list of impaired waters for exceeding the chloride standard for aquatic life. Shingle Creek is considered a single assessment reach for the purposes of evaluating compliance with State water quality standards. However, several water bodies are included in the Shingle Creek watershed that may have unique hydrologic conditions. This TMDL evaluates all stream reaches in the Shingle Creek watershed including Ryan Creek, Bass Creek, and Pike Creek in addition to Shingle Creek (Hydrologic Unit Code: 07010206-506).

2.2 APPLICABLE MINNESOTA WATER QUALITY STANDARDS AND ENDPOINTS

Shingle Creek is designated as Class 2 water for the protection of Aquatic Life (Minnesota R. ch. 7050). Chloride standards for the protection of these beneficial uses include a chronic standard of 230 mg/L based on the 4-day average and an acute standard of 860 mg/L for a one-hour duration for class 2 waters (Minnesota R. ch. 7050 and 7052).

2.3 MPCA NON-DEGRADATION POLICY

An important aspect of water quality standards in Minnesota is the non-degradation policy. The fundamental concept of non-degradation is the protection of water bodies already meeting State water quality standards. A more thorough discussion of Minnesota’s non-degradation policy can be found in MPCA’s “Guidance Manual for Assessing the Quality of Minnesota Surface Waters” (MPCA 2003). This TMDL was prepared in compliance with the State of Minnesota’s non-degradation policy.
3.0 Watershed Characterization

3.1 WATERSHED DESCRIPTION

The Shingle Creek watershed covers 44.5 square miles in east-central Hennepin County including nine municipalities (Figure 3.1). Shingle Creek begins at the junction of Bass Creek and Eagle Park in Brooklyn Park, flows easterly, then southerly for a total of 11.3 miles before discharging into the Mississippi River in Minneapolis. The nine municipalities included in the watershed are Brooklyn Center, Brooklyn Park, Crystal, Maple Grove, Minneapolis, New Hope, Osseo, Plymouth, and Robbinsdale. These entities created a joint powers organization, The Shingle Creek Watershed Management Commission (SCWMC), as required by the Metropolitan Surface Water Management Act of 1982. The SCWMC’s responsibilities include controlling excessive volumes and rate runoff, stormwater management, improving water quality, preventing flooding and erosion, promoting groundwater recharge, protecting and enhancing fish and wildlife habitat, and water recreation. In addition to these municipalities, roads in the watershed are also maintained by Hennepin County and the Minnesota Department of Transportation (Mn/DOT).
Watershed Base Map

Figure 3.1

Base data source: Minnesota Department of Natural Resources.
3.2 LAND USE

3.2.1 Current Land Use

Land use within the Shingle Creek and West Mississippi watershed has been and will be influenced by several factors, primarily proximity to Minneapolis and St. Paul and access to major transportation routes.

The predominant land uses in the southern and eastern part of the watershed are dense residential, commercial, and industrial, and in the northern and western part less dense residential, commercial, and industrial with some remaining undeveloped land (Figure 3.2; Table 3.1). All of the SCWMC except a small portion of the southwest corner of the watershed in Plymouth is within the existing Metropolitan Urban Service Area (MUSA). As such, metropolitan services and facilities including sanitary sewer are provided. Of that area of Plymouth in the SCWMC currently outside the MUSA, most lies within the MUSA 2020 expansion area. Plymouth has committed to protecting wetlands, lakes, and other natural resources within that expansion area as it develops.

Table 3.1. Land Use in the Shingle Creek Watershed

<table>
<thead>
<tr>
<th>Landuse</th>
<th>Area (acres)</th>
<th>Percent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Single Family Residential</td>
<td>8,759</td>
<td>30%</td>
</tr>
<tr>
<td>Roads and Major Highway</td>
<td>5,205</td>
<td>18%</td>
</tr>
<tr>
<td>Park, Recreational or Preserve</td>
<td>2,486</td>
<td>9%</td>
</tr>
<tr>
<td>Undeveloped</td>
<td>2,353</td>
<td>8%</td>
</tr>
<tr>
<td>Industrial and Utility</td>
<td>2,184</td>
<td>8%</td>
</tr>
<tr>
<td>Multi-Family Residential</td>
<td>1,696</td>
<td>6%</td>
</tr>
<tr>
<td>Commercial</td>
<td>1,507</td>
<td>5%</td>
</tr>
<tr>
<td>Institutional</td>
<td>1,290</td>
<td>4%</td>
</tr>
<tr>
<td>Water</td>
<td>1,271</td>
<td>4%</td>
</tr>
<tr>
<td>Extractive</td>
<td>1,183</td>
<td>4%</td>
</tr>
<tr>
<td>Airport</td>
<td>370</td>
<td>1%</td>
</tr>
<tr>
<td>Agriculture</td>
<td>285</td>
<td>1%</td>
</tr>
<tr>
<td>Mixed Use</td>
<td>94</td>
<td>0.3%</td>
</tr>
<tr>
<td>Railway</td>
<td>72</td>
<td>0.3%</td>
</tr>
<tr>
<td>Farmsteads</td>
<td>16</td>
<td>0.1%</td>
</tr>
<tr>
<td>TOTAL</td>
<td>28,771</td>
<td>100%</td>
</tr>
</tbody>
</table>
3.2.2 Population Density

In general, the central and southeastern part of the watersheds is developed, with population density increasing to the southeast. Minneapolis within the watershed is very dense, as are portions of adjacent Robbinsdale, Brooklyn Center, and Brooklyn Park. Significant areas of commercial/industrial development cluster around major highways: TH 100, TH 169, CSAH 81, I-94.

Only three significant undeveloped or lightly developed areas of the watershed remain: northern Brooklyn Park north of 85th Avenue, now quickly developing; in Maple Grove, the area around and including part of the gravel pits, being developed as the large Arbor Lakes multi-use development; and significant tracts in northwestern Plymouth. Development will intensify in some parts of Plymouth that are currently developed at a low density. However, significant tracts that are now undeveloped or developed at very low density are intended to remain that way.

3.2.3 Future Land Use

Areas of projected urban growth are shown in Figure 3.3. These data were compiled by the Metropolitan Council from cities’ most recent Comprehensive Plans, and represents cities’ expected 2020 land use. Most of the currently undeveloped or lightly developed areas of northern Brooklyn Park, southeastern Maple Grove, and northwestern Plymouth are shown as expected to be developed by 2020. Growth is expected to be a mix of development at different densities, and to include residential, commercial, and industrial uses.
2020 Land Use

- Single Family Residential
- Multi-Family Residential
- Commercial
- Industrial
- Institutional
- Transportation
- Park and Open Space
- Agricultural
- Mixed Use
- Vacant or No Data
- Open Water

Base data source: Minnesota Department of Natural Resources. Land use: Metropolitan Council from city comprehensive plans.

SHINGLE CREEK WATERSHED MANAGEMENT COMMISSION

2020 Land Use

Figure 3.3
3.3 SOILS

Most of the watersheds’ area is composed of well-drained soils. Texture is generally sandy or loamy with scattered organic or marsh soils areas. Highly to moderately permeable soils dominate the watershed, as indicated by large areas covered by soil hydrologic groups A and B. In poor permeability areas, soils are heavy textured soil groups such as clays/clay-loams and silt/silt-loams. Heavier soils can often result in reduced permeability.

3.4 GEOLOGY AND GEOMORPHOLOGY

Two major geomorphic regions are found in the Shingle Creek watershed: the Mississippi Valley Outwash area and the Emmons-Faribault moraine area. The outwash area is predominant in the eastern portion of the watersheds. The western portion of the watersheds is within the Emmons-Faribault moraine. This morainic area is characterized by a rolling topography with a relief of 20 to 30 feet. There are several lakes within this geomorphic area.

The surficial geology of the western half the watersheds ranges from areas of lacustrine sand and silt and clay and silt in the south to the sandy and loamy till in the north that characterizes the northwestern part of the county. Significant deposits of sand and gravel in the northwestern part of the watersheds are apparent in the gravel mining area of Maple Grove.

3.5 HYDROGRAPHIC DATA

Average daily flows have been monitored and reported at the USGS station at Queen Avenue since 1996. Additionally, stream flow was monitored at the outlet (Humboldt Avenue) and Zane Avenue by the SCWMC. Monthly average flows at the USGS station range from 2.77 cfs in January to 38 cfs in May. The maximum average daily flow at the USGS station was 225 cfs recorded on July 1, 1997.
3.6 METEOROLOGICAL DATA

Precipitation in the Twin Cities metropolitan area averages approximately 29 inches annually with average annual snowfall of 56 inches (State Climatology Office – Department of Natural Resources December 2000).

Chloride and discharge monitoring for the TMDL occurred from December 2002 through August 31, 2003. The winter of 2002-2003 was relatively mild with snowfall total of 36 inches (Table 3.2). However, data was collected by the USGS at the Queen Avenue Bridge from May of 1996 to December of 1998. The winter of 1996-1997 was a heavy snow year with 72.1 inches of snowfall. The winter of 1997-1998 was slightly below the average snowfall of 56 inches at 45 inches. These data were analyzed to address annual variability.

<table>
<thead>
<tr>
<th>Month</th>
<th>Snowfall (inches)</th>
<th>Twin Cities Area Precipitation or Water Equivalence (inches)</th>
<th>Difference from Normal(^1) (inches)</th>
</tr>
</thead>
<tbody>
<tr>
<td>September-2002</td>
<td>0</td>
<td>3.69</td>
<td>1.00</td>
</tr>
<tr>
<td>October-2002</td>
<td>0</td>
<td>3.80</td>
<td>1.69</td>
</tr>
<tr>
<td>November-2002</td>
<td>1.4</td>
<td>0.07</td>
<td>(1.87)</td>
</tr>
<tr>
<td>December-2002</td>
<td>3.0</td>
<td>0.28</td>
<td>(0.72)</td>
</tr>
<tr>
<td>January-2003</td>
<td>5.1</td>
<td>0.29</td>
<td>(0.75)</td>
</tr>
<tr>
<td>February-2003</td>
<td>10.7</td>
<td>0.81</td>
<td>0.02</td>
</tr>
<tr>
<td>March-2003</td>
<td>13.2</td>
<td>1.56</td>
<td>(0.30)</td>
</tr>
<tr>
<td>April-2003</td>
<td>1</td>
<td>2.61</td>
<td>0.30</td>
</tr>
<tr>
<td>May-2003</td>
<td>0</td>
<td>5.43</td>
<td>2.19</td>
</tr>
<tr>
<td>June-2003</td>
<td>0</td>
<td>3.57</td>
<td>(0.77)</td>
</tr>
<tr>
<td>July-2003</td>
<td>0</td>
<td>3.24</td>
<td>(0.80)</td>
</tr>
<tr>
<td>August-2003</td>
<td>0</td>
<td>0.69</td>
<td>(3.36)</td>
</tr>
<tr>
<td><strong>Total</strong></td>
<td><strong>34.4</strong></td>
<td><strong>26</strong></td>
<td><strong>(3.37)</strong></td>
</tr>
</tbody>
</table>

\(^1\)Values in parentheses are below normal

Snow pack loss and subsequent runoff is an important process in controlling chloride movement to surface waters. Maximum daily temperatures, snow pack depth, and discharge for the TMDL monitoring period are presented in Figure 3.4.
Figure 3.4. Maximum daily temperature, snow pack depth, and discharge in the Shingle Creek watershed for the winter of 2002-2003. Weather data was collected by the National Weather Service in New Hope.
Warm periods in the winter can result in melting of surface snow and increasing the snow water equivalence of the current snow pack and/or can result in a runoff event in the watershed. In general, late January and early February demonstrated an increase in snow pack depth. Following this period, snow pack depth decreased without significant runoff until about mid-February when a runoff event was recorded. This pattern demonstrates a period of snowmelt without runoff that increases the snow water equivalence.
4.0 Water Quality Monitoring Methods

In order to develop an understanding of chloride dynamics in an urban environment, monitoring of conductivity, chloride and discharge was performed from late November 2002 through August of 2003. All monitoring activities were outlined in a monitoring plan approved by the Technical Advisory Committee and MPCA (MWH, 2002). Following is a description of these activities and subsequent data processing.

4.1 STREAM SAMPLING LOCATIONS

Table 4.1 has a description of each of the stream monitoring locations. All of the sites are presented on Figure 4.1.

4.2 STREAM DISCHARGE AND CONDUCTIVITY MONITORING

Seven sites were continuously monitored for flow and conductivity (Figure 4). All sampling protocols followed an approved sampling plan (MWH 2001). Sampled was conducted from November 2002 through October of 2003. Grab samples for chloride were collected during base flow and runoff conditions at these sites to develop relationships between chloride and conductivity. Conductivity and stage were recorded every 15 minutes, and chloride samples collected biweekly and during significant runoff events. One sampling site was a storm sewer outfall that drains portions of Maple Grove. However, due to low flows, these data are not utilized in this analysis.
Table 4.1. Stream Sampling Sites in the Shingle Creek Watershed. River Mile (RM) is given for each site.

<table>
<thead>
<tr>
<th>Site Name</th>
<th>Stream Location Description</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Continuous Conductivity and Flow Monitoring Sites</strong></td>
<td></td>
</tr>
<tr>
<td>SC00</td>
<td>Shingle Creek outlet long term monitoring station.</td>
</tr>
<tr>
<td>SCI94</td>
<td>Shingle Creek downstream of I-94/694 Bridge</td>
</tr>
<tr>
<td>SC03</td>
<td>Shingle Creek Zane Avenue long term monitoring station.</td>
</tr>
<tr>
<td>SCSS2</td>
<td>A 60” concrete stormsewer pipe that drains to Shingle Creek. Automated conductivity measured at a manhole located just south of North Hennepin Community College and between Broadway and adjacent trail</td>
</tr>
<tr>
<td>SC04</td>
<td>Shingle Creek at east end of Northland Ct (The Quadrant office complex.) Sampling location is downstream of large wetland/stormwater pond.</td>
</tr>
<tr>
<td>SCSS1</td>
<td>Several stormsewers discharge to Bass Creek upstream of sampling location but station is below mixing zone</td>
</tr>
<tr>
<td>SCPINE</td>
<td>Upstream of Pineview and approximately 2000’ upstream of Bass Lake</td>
</tr>
<tr>
<td><strong>Grab Sample Sites</strong></td>
<td></td>
</tr>
<tr>
<td>Twin Lake Inlet</td>
<td>Ryan Creek</td>
</tr>
<tr>
<td>France Ave N</td>
<td>A low flow stream downstream of France between Twin Lake lower basin and Ryan Lake.</td>
</tr>
<tr>
<td>France</td>
<td>Inlet to Twin Lake upper basin: Upstream of Bass Lake Rd as it curves around Twin Lakes upper basin.</td>
</tr>
<tr>
<td>Xerxes Ave N</td>
<td>Shingle Creek downstream of Xerxes between 75th and Brookdale Dr. and adjacent to Palmer Lake Trail</td>
</tr>
<tr>
<td>62 East</td>
<td>Shingle Creek downstream of Hwy. 169 and upstream of large wetland complex between Hwy. 169 and Boone Ave N.</td>
</tr>
<tr>
<td>62 West</td>
<td>Pike Creek upstream of 62nd and approximately 1500’ upstream of Pike Lake</td>
</tr>
</tbody>
</table>

4.2.1 Stage Measurements, Rating Curves, and Discharge

Stage was monitored at four sites using SOLINST level loggers (pressure transducers). Data was collected at 15-minute intervals from late March through October 31, 2003. These data were adjusted to match a benchmark in the stream and corrected for barometric pressure. Details of the adjustments are documented in Appendix A. Stage data at Zane Ave. (SC03) and the Outlet (SC00) were collected using ISCO transducers. Stage-discharge rating curves were developed for each site. Details of rating curve development are in Appendix A.
4.2.2 Data Gaps

Although 15-minute stage data were collected at each of the monitoring sites in the watershed, there are periods where data could not be collected due to winter freeze potential (or where logger failure occurred). These data gaps were filled using regression equations relating the site with the long term USGS station at Queen Avenue. Two equations were used to fill data gaps. Summer and fall data were used to estimate winter discharge since these data are most representative of low flow periods. Spring equations were run separately since discharge in the spring is highly variable. Regression statistics are presented in Table 4.2.

<table>
<thead>
<tr>
<th>Site</th>
<th>Season</th>
<th>Slope</th>
</tr>
</thead>
<tbody>
<tr>
<td>62 East</td>
<td>Winter/Summer/Fall</td>
<td>0.298</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>0.234</td>
</tr>
<tr>
<td>SCI94</td>
<td>Winter/Summer/Fall</td>
<td>0.896</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>0.839</td>
</tr>
<tr>
<td>SC04</td>
<td>Winter/Summer/Fall</td>
<td>0.735</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>0.54</td>
</tr>
<tr>
<td>SCPINE</td>
<td>Winter/Summer/Fall</td>
<td>0.208</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>0.179</td>
</tr>
<tr>
<td>SC00</td>
<td>Winter/Summer/Fall</td>
<td>1.17</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>1.15</td>
</tr>
<tr>
<td>SC03</td>
<td>Winter/Summer/Fall</td>
<td>0.673</td>
</tr>
<tr>
<td></td>
<td>Spring</td>
<td>0.883</td>
</tr>
</tbody>
</table>

4.2.3 Winter Flow Estimates

Flow in the winter is difficult to estimate due to ice conditions and equipment limitations. However, winter flow is important to understanding chloride dynamics in the winter season. Winter flow estimates were generated using the seasonal regressions described in Section 4.2.2. However, it is important to note that winter stage was measured by the USGS using a pressure transducer at the Queen Avenue location. Stage measurements from pressure transducers can be susceptible to backwater effects caused by ice on the stream and can produce some sampling error in the calculated discharge. Spot-checking the data with loss of snow pack suggests that the results provide a good approximation of runoff events in the watershed. Winter flow was compared to changes in conductivity to further verify events. Since load analysis compares loads at the same flow point, comparisons during the winter month are not sensitive to these flow...
errors, rather are dependent upon robust concentration estimates. Further examination of winter flows was accomplished using the XP-SWMM hydraulic model.

4.3 GRAB SAMPLES

Samples were collected biweekly and during runoff events. All sampling protocols followed an approved sampling plan (MWH 2002). Sampling was conducted from November 2002 through August of 2003. Grab sampling occurred at all continuous and grab sample sites and included field measurements of conductivity, dissolved oxygen, and temperature.

4.4 ROAD SALT APPLICATION

Another key component of the field study was documentation of salt applied for deicing purposes. GIS was used to accurately quantify road salt applied to the watershed spatially and under varied intensities. The GIS data processing is briefly described in the following sections.

4.4.1 Road Surface Evaluation

The first step in the evaluation of road surfaces was to “burn” or introduce the road surfaces into the land use coverage. Existing land use coverages do not account for road areas except for a few major right-of-ways, representing roads with an over-laid line coverage that ignores road width. To estimate road width to add to the land use coverage, twenty-seven places were chosen to measure the width of the road, including shoulders, and ramps over the Metropolitan Council 2000 1-meter digital orthophotos for the Shingle Creek Watershed. These widths were used to determine the road areas from the Minnesota Department of Transportation (Mn/DOT) alignments and DOT Basemap Roads for Hennepin County (2001 GIS data). The remaining land uses were then reduced by the corresponding area converted to roadway. The base land use coverage is from the Metropolitan Council, and is representative of the generalized land use for the year 2000. Completion of this analysis resulted in a land use coverage with actual road areas
instead of lines representing roads of many different sizes. More details on this analysis can be found in Appendix B.

4.4.2 Salt Applied for Deicing

Agencies responsible for road deicing maintained records of salt applied for the winter of 2002 and 2003. All roads in the watershed were assigned one of three plow route types (Mn DOT, Hennepin County, or Municipality.) Municipality plow routes were specified by the cities in the watershed (Brooklyn Center, Brooklyn Park, Plymouth, Osseo, Robbinsdale, New Hope, Maple Grove, Crystal, and Minneapolis.) The lane miles were tabulated for each subwatershed by plow route type. The salt application data, in units of tons of salt applied per lane mile, coupled with the lane mile estimates were used to estimate the amount of salt applied to each subwatershed. For example, one subwatershed may cross three plow routes from three different applicators. Each of the applicators applies salt at a different rate for each event. The calculation assumes that in any given event, the driver is using the same application rate across the subwatershed boundaries. For example, if a driver reports using a total of 100 tons of salt for a 0.5 inch snowfall event, we assume that salt was applied evenly throughout that drivers route. Although there might be small variations in rates throughout the route, this approach provides a reasonable representation of where the salt ends up in the watershed. However, the rate is variable by event and is calculated from the reported application data provided by the drivers. - All of these records were compiled for the plow routes designated by the corresponding agency. Salt application records were then allocated to the appropriate subwatersheds using GIS on a daily time step.

NOTE: Mn/DOT uses Salt Institute research to create guidelines for Mn/DOT supervisors to determine the rates of salt application (varying between 100 to 800 lbs/mile). Mn/DOT supervisors analyze the information collected by the State’s Road Weather Information Systems (RWIS) and other sources to determine the rate of salt application that operators should use in the field. This rate guideline can also be altered by operators based on road conditions observed in the field.
4.5 SALT PILES AND RUNOFF

Salt piles in the watershed were inventoried and a site evaluation completed for each site. Site evaluations included assessment of storage area, drainage from the site, and general site information such as ground surface (i.e., gravel versus pavement). Salt piles were sampled for salt pile chemical composition. Ten representative samples from various places in the salt pile were collected with a stainless steel scoop and composited in a glass container collecting approximately one kilogram. These samples were analyzed for total and orthophosphorus. Additionally, two events were sampled from several of the sites to characterize salt pile runoff quality. Water samples were analyzed for chloride, total cyanide, free cyanide (HCN), total phosphorus, and orthophosphorus.

4.6 QUALITY CONTROL

Quality control is an important aspect of any sampling effort. Several measures were in place during the filed investigations including collecting duplicate samples and calibration analysis of field loggers.

4.6.1 Grab Samples

Twenty duplicate samples were taken representing 9% of the total samples collected. There was generally a less than 10% difference between duplicate samples collected during the field study (Figure 4.2).
4.6.2 Conductivity Loggers

Conductivity loggers were checked using both standards and an independent field conductivity meter. Conductivity loggers were evaluated and calibrated once each in April, July, and October by comparing the measured conductivity in a standard to the standard value. Evaluation of the loggers demonstrates that measurements were typically within 10% of conductivity standards with a few exceptions. The conductivity loggers performed very well.

Logged conductivity was also compared to an independent field measure of conductivity (Figure 4.3). With one exception, field and logged conductivity were typically within 10% with the median difference of less than 3%.
Figure 4.3. Logged and Field Measured Conductivity Plotted along a 1:1 Line.
5.0 Source Assessment

Chloride can originate from a wide range of sources including industrial wastewater discharge, municipal wastewater treatment plant effluent, runoff from road application of salt for deicing, runoff from parking lots and fertilizer applications. A detailed assessment of sources in the Shingle Creek watershed was conducted as a part of this TMDL.

5.1 POINT SOURCES

There are few point sources in the Shingle Creek watershed. There are no wastewater treatment plant effluent discharges in the watershed. NPDES permits in the watershed are listed in Table 5.1. None of the SC permits attached have chloride as a parameter of concern (Nancy Drach, MPCA pers. comm.). Consequently, the NPDES permit holders listed in Table 5.1 are all considered deminimus in regard to chloride discharges. Therefore, these discharges are considered insignificant sources and are not assigned a waste load allocation in this TMDL. The Hutchinson Technology permit lists coolant water as treated by reverse osmosis as being discharged.

<table>
<thead>
<tr>
<th>NPDES ID</th>
<th>Facility Name</th>
<th>Address</th>
<th>SIC Description</th>
</tr>
</thead>
<tbody>
<tr>
<td>MNG490009</td>
<td>C S McCrossan</td>
<td>7865 Jefferson Hwy Maple Grove</td>
<td>Asphalt Paving Mixtures and Blocks</td>
</tr>
<tr>
<td>MNG250048</td>
<td>Robinson Rubber Products Co Inc</td>
<td>4600 Quebec Ave N New Hope</td>
<td>Fabricated Rubber Products</td>
</tr>
<tr>
<td>MN0002119</td>
<td>GAF Materials</td>
<td>49th Avenue Minneapolis</td>
<td>Asphalt Felts and Coatings</td>
</tr>
<tr>
<td>MNG490010</td>
<td>Tiller Corp</td>
<td>10633 89th Ave N Maple Grove</td>
<td>Asphalt Paving Mixtures and Blocks</td>
</tr>
<tr>
<td>MNG790069</td>
<td>Former TPI Facility - 9145</td>
<td>6830 Brooklyn Blvd Brooklyn Center</td>
<td>Gasoline Service Station</td>
</tr>
<tr>
<td>MNU000378</td>
<td>Universal Foods</td>
<td>New Hope</td>
<td></td>
</tr>
<tr>
<td>MNU790130</td>
<td>Former Pilgrim Cleaners</td>
<td>Brooklyn Blvd &amp; 69th Brooklyn Center</td>
<td>Dry Cleaner</td>
</tr>
<tr>
<td>MN0066699</td>
<td>Hutchinson Technology</td>
<td>5905 Trenton Plymouth</td>
<td>Metal Stamping</td>
</tr>
<tr>
<td>MN0066958</td>
<td>Mn/DOT TH 100 Project</td>
<td>Robbinsdale &amp; Brooklyn Center</td>
<td>Highway Construction Dewatering</td>
</tr>
</tbody>
</table>
In addition to these NPDES permits in the watershed, NPDES Phase II permits for small municipal separate storm sewer systems (MS4) have been issued to the member cities in the watershed as well as Hennepin County and Mn/DOT. The City of Minneapolis has an individual NPDES permit for Stormwater – NPDES Permit # MN 0061018. The other cities, Hennepin County and MnDOT Metro District, are covered under the Phase II General NPDES Stormwater Permit – MNR040000. The unique permit numbers assigned to these cities, Hennepin County and MnDOT Metro District are as follows:

- Brooklyn Center – MS400006
- Brooklyn Park – MS400007
- Crystal – MS400012
- Maple Grove – MS400102
- New Hope – MS400039
- Osseo – MS400043
- Plymouth – MS400112
- Robbinsdale – MS400046
- Hennepin County – MS400138
- MnDOT Metro District – MS400170

EPA requires that stormwater discharges regulated under NPDES be allocated into the wasteload allocation or point source portion of the TMDL. Although the sources of chloride in the watershed are nonpoint in nature, they are allocated in the wasteload allocation in this TMDL. However, the discussion of the sources maintains the nonpoint source nature of chloride.

### 5.2 NON-POINT SOURCES

The majority of chloride in the Shingle Creek watershed is derived from nonpoint sources including road deicing, commercial and industrial deicing, and fertilizer application. Most fertilizer application occurs in the spring, summer, and fall suggesting that the chloride generated from this source either infiltrates into the groundwater or runs off during spring and summer storms.
5.2.1 Salt Piles

Salt piles are a potential source of chloride in the Single Creek watershed. Salt piles or road salt storage facilities are used to store road salt before application to roads for snow and ice removal. Table 5.2 lists the salt piles in the Shingle Creek watershed along with some general characteristics of the storage facility. There are eight salt piles in the Shingle Creek watershed.

Several factors can affect the amount of chloride that can enter stream systems from a road salt storage facility. In general, covered road salt piles with an impervious surface will generate less runoff and infiltration of chloride-laden water. Two of the salt piles in the watershed were only covered by a tarp and one of these was on a gravel surface. The drainage route can also affect the amount of chloride discharge to surface waters. Direct connections through storm pipes provide a direct route to surface waters whereas discharge to a pond can offer some retention and dilution of salt storage facility runoff. Most of the facilities drained to a pond or wetland and then directly to a storm sewer. Runoff chloride, phosphorus and cyanide concentrations were measured for several of these salt storage facilities.

<table>
<thead>
<tr>
<th>Operator</th>
<th>Location</th>
<th>Storage Facility</th>
<th>Pile Composition</th>
<th>Drainage Surface</th>
<th>Drainage Route</th>
</tr>
</thead>
<tbody>
<tr>
<td>Hennepin County</td>
<td>West of Hwy 81</td>
<td>Unknown</td>
<td>Unknown</td>
<td>Unknown</td>
<td>Unknown</td>
</tr>
<tr>
<td>Osseo</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Maple Grove</td>
<td>Forestview La. N.</td>
<td>Covered with plastic tarp on asphalt</td>
<td>Salt</td>
<td>Asphalt</td>
<td>Surface drainage to wetland 50 ft from pile; discharge from wetland to storm sewer</td>
</tr>
<tr>
<td>Brooklyn Park</td>
<td>Noble Ave. N. north of 83rd Ave N.</td>
<td>Enclosed</td>
<td>Salt</td>
<td>Asphalt</td>
<td>Surface drainage to pond 300 ft from pile; discharge from pond to storm sewer</td>
</tr>
<tr>
<td>Brooklyn Center</td>
<td>Shingle Creek Pkwy. east of Shingle Creek</td>
<td>Enclosed</td>
<td>Salt</td>
<td>Asphalt</td>
<td>Surface drainage to storm sewer to pond</td>
</tr>
<tr>
<td>Robbinsdale</td>
<td>Toledo Ave. north of 45th Ave. N.</td>
<td>Covered with plastic tarp on gravel</td>
<td>Salt/sand mixture</td>
<td>Gravel</td>
<td>Surface drainage to ditch adjacent to property; ditch drains to storm sewer</td>
</tr>
<tr>
<td>New Hope</td>
<td>International Pkwy. south of Research Center Rd. E.</td>
<td>Enclosed</td>
<td>Salt</td>
<td>Asphalt</td>
<td>Surface drainage to storm sewer</td>
</tr>
<tr>
<td>Osseo</td>
<td>Broadway Ave. west of Hwy. 169</td>
<td>Covered with plastic tarp on asphalt</td>
<td>Salt/sand mixture</td>
<td>Asphalt</td>
<td>Surface drainage to storm sewer</td>
</tr>
<tr>
<td>Crystal</td>
<td>41st Ave N. east of Douglas Dr. N.</td>
<td>Enclosed</td>
<td>Salt</td>
<td>Asphalt</td>
<td>Surface drainage to pond south of property</td>
</tr>
</tbody>
</table>
Spillage of road salt and deicing materials can also increase the amount of chloride in runoff from salt storage facilities. Spillage outside of covered areas makes the road salt available for dissolution and runoff during precipitation events.

Another potential source of chloride from road salt storage facilities is the washing of the maintenance vehicles. Wash water that enters the storm sewer system ultimately ends up in surface waters. Although this source is potentially small in comparison to other sources in the watershed, it is worth noting.

Runoff from salt piles in the watershed was sampled on March 20, March 28 and April 17, 2003. Samples were analyzed for ortho and total phosphorus as well as chloride and total and free cyanide (weak acid dissociable). Results of these sampling events are presented in Table 5.3.

<table>
<thead>
<tr>
<th>Operator Area (ac)</th>
<th>Drainage Route</th>
<th>Chloride (mg/L)</th>
<th>Free Cyanide (mg/L)</th>
<th>Total Cyanide (mg/L)</th>
<th>Total Phosphorus (mg/L)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Hennepin County Osseo 0.10</td>
<td>Unknown</td>
<td>1,270</td>
<td>ND</td>
<td>0.078</td>
<td>0.219</td>
</tr>
<tr>
<td>Maple Grove 0.07</td>
<td>Surface drainage to wetland 50 ft from pile; discharge from wetland to storm sewer</td>
<td>12,800</td>
<td>0.014</td>
<td>0.904</td>
<td>0.119</td>
</tr>
<tr>
<td>Brooklyn Park 0.27</td>
<td>Surface drainage to pond 300 ft from pile; discharge from pond to storm sewer</td>
<td>824</td>
<td>ND</td>
<td>0.103</td>
<td>0.175</td>
</tr>
<tr>
<td>Brooklyn Center 0.32</td>
<td>Surface drainage to storm sewer to pond</td>
<td>--</td>
<td>--</td>
<td>--</td>
<td>--</td>
</tr>
<tr>
<td>Robbinsdale 0.06</td>
<td>Surface drainage to ditch adjacent to property; ditch drains to storm sewer</td>
<td>1,038</td>
<td>ND</td>
<td>0.016</td>
<td>0.162</td>
</tr>
<tr>
<td>New Hope 0.16</td>
<td>Surface drainage to storm sewer</td>
<td>19</td>
<td>ND</td>
<td>ND</td>
<td>0.070</td>
</tr>
<tr>
<td>Osseo 0.05</td>
<td>Surface drainage to storm sewer</td>
<td>1,285</td>
<td>ND</td>
<td>0.037</td>
<td>0.257</td>
</tr>
<tr>
<td>Crystal 0.20</td>
<td>Surface drainage to pond south of property</td>
<td>17</td>
<td>ND</td>
<td>ND</td>
<td>0.137</td>
</tr>
</tbody>
</table>

### 5.2.2 Road Deicing

One of the primary sources of chloride in the watershed is the application of road salt or road salt alternatives in the watershed. The predominant chloride salt used for deicing in North America is sodium chloride (Environment Canada 1999). Substances potentially present in road salt include phosphorus (14-26 mg/kg), nitrogen (6.8-4,200 mg/kg), copper (0-14 mg/kg), and zinc (0.02 – 0.68 mg/kg) (MDOT 1993). Additives often include sodium ferrocyanide and ferric ferrocyanide used as anti-caking agents. These additives are of some concern because these
compounds can photolyse and release free cyanide ions which are toxic to aquatic organisms. Runoff concentrations from salt piles in the Shingle Creek watershed only found one detection of free cyanide (Table 5.3) and several grab samples collected from Shingle Creek were non-detects as well.

Table 5.4 presents results from salt pile sampling in the Shingle Creek watershed. Salt piles were sampled at 10 different locations vertically and then composited and analyzed for total and orthophosphorus. Total phosphorus concentrations ranged from 6.3 to 28 ppm.

Table 5.4. Phosphorus results from salt pile sampling for salt storage areas that supply salt for use in the Shingle Creek Watershed.

<table>
<thead>
<tr>
<th>Salt Pile</th>
<th>Ortho P (mg/kg)</th>
<th>Total P (mg/kg)</th>
</tr>
</thead>
<tbody>
<tr>
<td>MNDOT Golden Valley</td>
<td>4.24</td>
<td>6.33</td>
</tr>
<tr>
<td>MNDOT Maple Grove</td>
<td>--</td>
<td>--</td>
</tr>
<tr>
<td>Hennepin County Osseo</td>
<td>--</td>
<td>--</td>
</tr>
<tr>
<td>Plymouth</td>
<td>ND</td>
<td>ND</td>
</tr>
<tr>
<td>Maple Grove</td>
<td>ND</td>
<td>6.77</td>
</tr>
<tr>
<td>Brooklyn Park</td>
<td>ND</td>
<td>ND</td>
</tr>
<tr>
<td>Brooklyn Center</td>
<td>ND</td>
<td>ND</td>
</tr>
<tr>
<td>Robbinsdale</td>
<td>ND</td>
<td>28</td>
</tr>
<tr>
<td>New Hope</td>
<td>ND</td>
<td>19.5</td>
</tr>
<tr>
<td>Osseo</td>
<td>1.16</td>
<td>13.4</td>
</tr>
<tr>
<td>Crystal</td>
<td>ND</td>
<td>ND</td>
</tr>
</tbody>
</table>

Roads in the Shingle Creek watershed are maintained by Mn/DOT, Hennepin County and the respective cities (Table 5.5). Hennepin County and Brooklyn Park maintain the largest proportion of roads comprising 37% of all the lane miles in the watershed.

Table 5.5. Lane Miles by Maintenance Official in the Shingle Creek Watershed.

<table>
<thead>
<tr>
<th>Owner</th>
<th>Lane Miles</th>
<th>Percent</th>
</tr>
</thead>
<tbody>
<tr>
<td>Hennepin County</td>
<td>259.9</td>
<td>19%</td>
</tr>
<tr>
<td>Brooklyn Park</td>
<td>243.2</td>
<td>18%</td>
</tr>
<tr>
<td>Mn DOT</td>
<td>155.9</td>
<td>11%</td>
</tr>
<tr>
<td>Brooklyn Center</td>
<td>139.1</td>
<td>10%</td>
</tr>
<tr>
<td>Crystal</td>
<td>112.0</td>
<td>8%</td>
</tr>
<tr>
<td>Minneapolis</td>
<td>105.7</td>
<td>8%</td>
</tr>
<tr>
<td>Plymouth</td>
<td>92.8</td>
<td>7%</td>
</tr>
<tr>
<td>Robbinsdale</td>
<td>87.8</td>
<td>6%</td>
</tr>
<tr>
<td>Maple Grove</td>
<td>86.9</td>
<td>6%</td>
</tr>
<tr>
<td>New Hope</td>
<td>73.5</td>
<td>5%</td>
</tr>
<tr>
<td>Osseo</td>
<td>18.7</td>
<td>1%</td>
</tr>
<tr>
<td>Total</td>
<td>1375.5</td>
<td>100%</td>
</tr>
</tbody>
</table>
Road salt applied in the watershed was typically sodium chloride applied in rock or brine form, often as a part of a mixture of salt and sand (Table 5.6).

<table>
<thead>
<tr>
<th>Road Authority</th>
<th>De-icing Substances Used</th>
<th>Comments</th>
</tr>
</thead>
<tbody>
<tr>
<td>Brooklyn Center</td>
<td>100% salt</td>
<td>Salt-sand used as necessary</td>
</tr>
<tr>
<td>Brooklyn Park</td>
<td>100% salt</td>
<td>Salt-sand used as necessary</td>
</tr>
<tr>
<td>Crystal</td>
<td>4:1 sand/salt</td>
<td></td>
</tr>
<tr>
<td>Maple Grove</td>
<td>100% salt</td>
<td>Have tried molasses product in past but had trouble with application – too sticky</td>
</tr>
<tr>
<td>Minneapolis</td>
<td>100% salt</td>
<td></td>
</tr>
<tr>
<td></td>
<td>5:1 sand/salt</td>
<td></td>
</tr>
<tr>
<td>New Hope</td>
<td>2:1 sand/salt</td>
<td></td>
</tr>
<tr>
<td>Osseo</td>
<td>1:1 sand/salt</td>
<td>Had good luck with “Clear Lane” MgCl/molasses product instead of salt in 2003-04 and will likely continue in the future</td>
</tr>
<tr>
<td>Plymouth</td>
<td>3:1 sand/salt</td>
<td>Occasional 100% salt</td>
</tr>
<tr>
<td>Robbinsdale</td>
<td>4:1 sand/salt</td>
<td></td>
</tr>
<tr>
<td>Hennepin County</td>
<td>100% Salt Sand/salt mix</td>
<td>5:1, 10:1 salt/sand as necessary. Has tried prewetting with mixed results. Have a potassium acetate test site outside of SC watershed on CR 135.</td>
</tr>
<tr>
<td>Mn/DOT</td>
<td>100% salt Sand/salt mix Some calcium chloride and magnesium chloride</td>
<td>Salt/sand of various mixes used as necessary.</td>
</tr>
<tr>
<td>CP Railroad Yard</td>
<td>Some sand/salt mix on rails and walkways as necessary</td>
<td>Some CaCl used in Feb-Mar to deice and dry out</td>
</tr>
</tbody>
</table>

Road salt application rates in the winter of 2002 and 2003 by maintenance entity is presented in Figure 5.1. Application rates were normalized to present rates in tons applied per lane mile by month and entity. Application occurs on some major highway shoulders to provide access for busses and mass transit. These lane miles were not included in these calculations. Application rates varied by maintenance entity, with the highest application rates associated with those entities responsible for major highways.
Figure 5.1. Road Salt Application Rates for each Month of the 2002-2003 Winter Season.

Approximately 8,701 tons of road salt (5,308 tons chloride) were applied to the watershed during the winter of 2002 and 2003 (Table 5.7). The heaviest application occurred in January and February, corresponding to the months with the greatest amount of snowfall. It is important to note that the winter of 2002-2003 was a below normal snow fall year for the Twin Cities metropolitan area. Snowfall was around 36 inches while the long-term average is approximately 56 inches. Data is not available for specific application amounts in the Shingle Creek watershed for other years. Consequently, we must assume that the rates in the monitored year are indicative of relative agency application rates. Stream data is available for 1996 through 1998 from the USGS and is used to assess interannual variability.

Table 5.7. Tons of Road Salt and Associated Chloride applied to the Shingle Creek Watershed during the Winter of 2002-2003 for Road Deicing.

<table>
<thead>
<tr>
<th>Month</th>
<th>Total Road Salt (tons)</th>
<th>Total Chloride (tons)</th>
</tr>
</thead>
<tbody>
<tr>
<td>October</td>
<td>1</td>
<td>0</td>
</tr>
<tr>
<td>November</td>
<td>6</td>
<td>3</td>
</tr>
<tr>
<td>December</td>
<td>773</td>
<td>471</td>
</tr>
<tr>
<td>January</td>
<td>3,414</td>
<td>2,083</td>
</tr>
<tr>
<td>February</td>
<td>2,360</td>
<td>1,440</td>
</tr>
<tr>
<td>March</td>
<td>2,026</td>
<td>1,236</td>
</tr>
<tr>
<td>April</td>
<td>122</td>
<td>75</td>
</tr>
<tr>
<td><strong>TOTAL</strong></td>
<td><strong>8,701</strong></td>
<td><strong>5,308</strong></td>
</tr>
</tbody>
</table>
5.2.3 Private Industrial and Residential Deicing

Private contractors, industry, and agencies such as port authorities and airports use salt as a deicer. Limited data were available for parking lots, industrial, commercial, and other private properties. Cheminfo (1999) estimated that commercial and industrial consumers represented approximately 5 to 10% of the road salt market. In quantifying total road salt application in Canada, Environment Canada used the midpoint of these data (7.5%) to represent commercial and industrial road salt application (Environment Canada 1999).

5.2.4 Natural Sources

Natural sources of chloride salts (calcium, potassium, sodium, and magnesium) can occur as a result of rock weathering, soil erosion, and atmospheric precipitation. Atmospheric precipitation is typically only important in coastal maritime regions. Local precipitation monitoring only identifies trace amounts of chloride in precipitation (NADP 2002). Few, if any, rock outcrops occur in the watershed. Consequently, any input from geologic sources would be groundwater sources.

5.2.5 Groundwater Discharge

Although groundwater sources are not directly addressed in this report, they can be important since much of what enters the groundwater can end up in the stream channel. Natural sources of chloride in groundwater are primarily geologic. Anthropogenic sources to groundwater can include septic leachate, landfill leachate, infiltration from fertilizers (potassium chloride), and infiltration of chloride rich runoff from deicing activities.

5.2.5.1 Water Softeners and Septic Systems

There is little information available for septic systems in the Shingle Creek watershed since most of the watershed is sewered. However, some septic systems do exist in the watershed. Typical chloride concentrations in untreated domestic wastewater range from 30 to 100 mg/L (Metcalf
and Eddy 1991). Much of this discharge would ultimately end up in groundwater through infiltration and not in surface waters.

Water softeners have also been mentioned as a potential source of chloride to surface waters. Concerns arise when the water softening system recharges resin with salt brine and discharge the wastewater rich in chloride. Most softened water is discharged to sanitary sewer systems and ultimately ends up in wastewater treatment plant effluent. Some may end up in septic systems. It is unlikely that this is a significant source in the Shingle Creek Watershed. Few septic systems exist in the watershed and there are no wastewater treatment plant discharges in the watershed. Of the septic systems that do exist, it is unclear as to the proportion that use water softeners.

5.2.5.2 Landfills

There are a few permitted and unpermitted landfills or dumps in the Shingle Creek watershed. Although these would be considered groundwater sources and are not addressed directly as a part of this TMDL, they are worth noting.

Several permitted and unpermitted solid waste and dumpsites are located in the Maple Grove Gravel Pits Area. Permitted sites include: North Hennepin Yard Waste site, Recycling Transfer Station, and Solid Waste Transfer Station. Unpermitted sites include: the Osseo/Maple Grove Pay Dump north of 85th and the Sonny Link Dump south of 85th, and an NSP fly ash dump between Jefferson Highway and TH 169, north of 83rd.

An unpermitted cement washings dump is on Shingle Creek south of Brooklyn Boulevard, west of CR 81. The old Brooklyn Park dump stood where Brooklyn Park Central Park is now located, south of 85th between Noble and Regent Avenues.

The old Brooklyn Center dump was located on 65th Avenue west of Brooklyn Boulevard.

More information can be found at:

http://pca-gis04.pca.state.mn.us/website/mes/mesfin/entry.htm
5.2.5.3 Fertilizers

Fertilizers used on lawns and landscaping often contain potassium chloride as a potassium source for plants. Consequently, fertilizers represent a potential source of chloride in the watershed. Much of the fertilizer would be applied in the spring, summer, and fall months to coincide with the growing season. Ultimately, chloride from fertilizers would enter surface waters as a result of runoff events soon after application or enter groundwater as a result of infiltration. Because of the timing of fertilizer application, it is unlikely that it represents a significant source during the most sensitive times for chloride (winter flow). The greatest potential for fertilizer chloride to reach surface waters is through ground water. Chloride from fertilizer application is considered a groundwater source in this TMDL.

5.2.5.4 Infiltration

Infiltration of surface water can also be a major source of chloride to groundwater. Infiltration water may be rich in chloride as a result of road application for deicing or fertilizer application.

5.2.6 Railway and Airport Deicing

Aviation activity at the Crystal Airport is sharply reduced in winter, and deicing of aircraft is not performed. Planes are typically grounded during inclement weather. Urea is used in a limited manner on runways in the winter with an estimated use less than 500 pounds per year. Some sand is used as an abrasive. However, no salt is used due to corrosive effects on aircraft.

The railways do apply a small amount of salt and sand, primarily to walkways in the Soo Line Humboldt switching yards. Some CaCl is used at the yards, primarily in February through March to deice and also to dry out the rail area. Salt, sand and CaCl are applied as needed and where needed, although there is no written or unwritten policy. There are no records of applications. Very little ice control is done in the rail corridor to the west. They do plow at the yards and the snow is stockpiled on site.
6.0 Assessment of Water Quality Data and Monitoring Results

6.1 HISTORIC DATA AND CAUSE FOR LISTING

The listing of Shingle Creek as impaired resulted from a limited sampling of chloride completed in 1996 by the US Geological Survey (USGS) at the Queen Avenue Bridge in Minneapolis. After reviewing the USGS data from Queen Avenue, the Shingle Creek WMO has been sampling routinely for chloride in Shingle Creek.

6.2 EXTENT OF CHLORIDE EXCEEDANCES

One of the primary goals of this TMDL was to determine the spatial extent, severity and duration of chloride exceedances in the Shingle Creek watershed. To define the extent of chloride exceedances in the watershed, both grab samples and logged conductivity data were collected at numerous sites throughout the watershed (Figure 4.1). Conductivity can act as a surrogate measure for chloride. Chloride is a charged ionic species that makes water conductive. As chloride concentrations increase, the conductivity of a solution increases; therefore, specific conductance and chloride are directly related. By utilizing conductivity as a surrogate for chloride and developing chloride-conductivity relationships, more robust data sets can be developed to increase the accuracy of load estimations and decrease the need for some manual data-collection activities. Additionally, the chronic standard is based on a four-day exposure to chloride concentrations. This is difficult to measure with grab samples unless data is collected daily. Logging specific conductance allows for the calculation of a four-day average to identify both the severity and duration of the exceedance.
6.2.1 Grab Samples

As expected, grab samples throughout the watershed demonstrated both chronic and acute exceedances. Stream grab sample concentrations ranged from 16 to 12,000 mg/L (Table 6.1). In box plots (Figure 6.1 and 6.2), the upper and lower ends of the box represent the 75\textsuperscript{th} and 25\textsuperscript{th} percentile while the line in the box represents the median value. Median values were higher at the three lowest sites in the watershed than the three higher sites. Bass Creek did not demonstrate any acute exceedances but the maximum of the grab samples did exceed the chronic standard.

Table 6.1. Grab Sample Results for the Shingle Creek Watershed.

<table>
<thead>
<tr>
<th>Creek</th>
<th>Site</th>
<th>N</th>
<th>Mean</th>
<th>Median</th>
<th>Min</th>
<th>Max</th>
</tr>
</thead>
<tbody>
<tr>
<td>Shingle Creek</td>
<td>SCSS1</td>
<td>17</td>
<td>793</td>
<td>180</td>
<td>55</td>
<td>8,200</td>
</tr>
<tr>
<td></td>
<td>SC04</td>
<td>18</td>
<td>180</td>
<td>125</td>
<td>66</td>
<td>700</td>
</tr>
<tr>
<td></td>
<td>SC03</td>
<td>27</td>
<td>308</td>
<td>150</td>
<td>16</td>
<td>2,900</td>
</tr>
<tr>
<td></td>
<td>Xerxes</td>
<td>19</td>
<td>297</td>
<td>210</td>
<td>68</td>
<td>1,200</td>
</tr>
<tr>
<td></td>
<td>SCI94</td>
<td>15</td>
<td>224</td>
<td>200</td>
<td>64</td>
<td>570</td>
</tr>
<tr>
<td></td>
<td>SC00</td>
<td>30</td>
<td>297</td>
<td>170</td>
<td>68</td>
<td>2,200</td>
</tr>
<tr>
<td>Bass Creek</td>
<td>SCPINE</td>
<td>13</td>
<td>120</td>
<td>100</td>
<td>33</td>
<td>420</td>
</tr>
<tr>
<td>Ryan Creek</td>
<td>Twin Lake</td>
<td>13</td>
<td>1069</td>
<td>150</td>
<td>64</td>
<td>12,000</td>
</tr>
<tr>
<td></td>
<td>France</td>
<td>15</td>
<td>575</td>
<td>84</td>
<td>51</td>
<td>3,400</td>
</tr>
<tr>
<td></td>
<td>Russell</td>
<td>6</td>
<td>85</td>
<td>76</td>
<td>35</td>
<td>170</td>
</tr>
<tr>
<td>Pike Creek</td>
<td>169</td>
<td>17</td>
<td>111</td>
<td>87</td>
<td>74</td>
<td>350</td>
</tr>
<tr>
<td></td>
<td>62 West</td>
<td>17</td>
<td>1031</td>
<td>260</td>
<td>67</td>
<td>7,400</td>
</tr>
<tr>
<td>Storm Sewer</td>
<td>SCSS2</td>
<td>16</td>
<td>3197</td>
<td>205</td>
<td>14</td>
<td>35,000</td>
</tr>
</tbody>
</table>

Figure 6.1. Box Plot of Grab Samples Collected from Shingle Creek
6.2.2 Chloride and Conductivity Relationships

Specific conductance was logged at a 15-minute interval at six sites in the watershed. At each of these sites, grab samples were also collected for chloride to develop a relationship between specific conductance and chloride concentrations for each site. Conductivity-chloride relationships are presented in Figures 6.3a and 6.3b. For all of the regression equations, the intercepts were forced through zero so that no negative values would be predicted. This stands to reason since natural streams in Minnesota would have some chloride and zero conductance would relate to water with no dissolved solids including chloride. These relationships were used to predict daily chloride concentrations at these sites.

Thorough examination of the regressions resulted in the identification of a few trends that need to be addressed, the first of which was the examination of the effects of outliers. Several extreme measurements occurred during the development of the relationships Table 6.2. Extreme values can have a disproportionate effect of a regression relationship causing an over or under prediction of the predicted variable. Since our analyses focuses on values around the standard concentrations of 230 mg/L and 860 mg/L, these extreme values were excluded from the relationships used to predict chloride concentrations.
Table 6.2. Extreme Conductivity and Chloride Values

<table>
<thead>
<tr>
<th>River Mile</th>
<th>Site</th>
<th>Conductivity (µS/cm)</th>
<th>Chloride (mg/L)</th>
</tr>
</thead>
<tbody>
<tr>
<td>0.6</td>
<td>SC00</td>
<td>8,210</td>
<td>2,200</td>
</tr>
<tr>
<td>7.3</td>
<td>SC03</td>
<td>8,255</td>
<td>2,900</td>
</tr>
<tr>
<td>11.4</td>
<td>SCSS1</td>
<td>26,800</td>
<td>8,200</td>
</tr>
<tr>
<td>11.4</td>
<td>SCSS1</td>
<td>5,750</td>
<td>1,900</td>
</tr>
</tbody>
</table>

Secondly, relationships between chloride and conductivity were examined seasonally to evaluate potential differences in the relationship that may result from changes in the proportion of the total dissolved solids represented by chloride. Our results indicate that winter runoff conductivity is most likely driven by deicing salt high in chloride whereas total dissolved solids in groundwater that may have proportionally less chloride contributing to the ionic balance may drive summer low flow conductivity. Once the outliers were removed and the seasonal variations taken into account, the relationships for the winter/spring period and summer period were significantly different with a summer slope for each of the sites around 0.15 and winter/spring slope around 0.21 in Table 6.3. The only exception was the Bass Creek site (Pineview; RM 14) where some of the weakest relationships occurred. It may be that this site is affected by groundwater during a greater portion of the year.

Table 6.3. Conductivity – Chloride Relationships in the Shingle Creek Watershed

<table>
<thead>
<tr>
<th>River Mile</th>
<th>Site</th>
<th>Summer</th>
<th>Winter/Spring</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>Slope</td>
<td>r-square</td>
</tr>
<tr>
<td>0.6</td>
<td>SC00</td>
<td>0.15</td>
<td>0.86</td>
</tr>
<tr>
<td>3.3</td>
<td>SCI94</td>
<td>0.14</td>
<td>0.99</td>
</tr>
<tr>
<td>7.3</td>
<td>SC03</td>
<td>0.16</td>
<td>0.81</td>
</tr>
<tr>
<td>10.3</td>
<td>SC04</td>
<td>0.16</td>
<td>0.9</td>
</tr>
<tr>
<td>11.4</td>
<td>SCSS1</td>
<td>0.15</td>
<td>0.97</td>
</tr>
<tr>
<td>14</td>
<td>Pineview</td>
<td>0.09</td>
<td>0.91</td>
</tr>
<tr>
<td></td>
<td>Standard Deviation</td>
<td>0.025</td>
<td>--</td>
</tr>
</tbody>
</table>

The slope values in Table 6.3 were used to predict chloride concentrations for each of the sites. Since there were only three points on the summer relationship at all of the sites except RM 0.6 and 7.3, predicted summer concentrations were based on the combined relationship of all of these sites combined (slope =0.15) except for RM 14.
Figure 6.3a. Chloride-Conductivity Relationships for Samples Collected in the Winter and Spring of 2002-03.
Figure 6.3b. Chloride-Conductivity Relationships for Samples Collected in the Summer of 2002-03.
The USGS also collected data at the Queen Avenue Bridge from May of 1996 to December of 1998. These data were used to develop chloride-conductivity relationships for the Queen Avenue site. After separating the data into winter and spring/summer/fall sets and forcing the intercept through zero, the slope values align with the data previously presented (Figure 6.4).

Figure 6.4. Chloride Conductivity relationship at the Queen Avenue Bridge. Data was collected by the USGS. The triangles represent summer/spring/fall data and the squares represent winter data.

### 6.2.2 Conductivity and Chloride Time Series

Time series were generated for chloride concentrations based on the logged conductivity. Two series were generated. The first was four-day average chloride concentrations with flow and grab chloride samples included on the plots. The second set of plots includes the daily maximum chloride concentration to assess acute exceedances.
6.2.2.1 Chronic Exceedances

A box plot of chloride concentrations based on measured conductivity by river mile is presented in Figure 6.5.

![Figure 6.5. Box Plot of Conductivity Estimated Chloride Concentrations in the Shingle Creek Watershed](image)

Figure 6.5 presents four-day average chloride concentrations based on the chloride conductivity relationships at six sites in the Shingle Creek watershed. All of the sites demonstrated exceedances during the winter months. Concentrations at River Mile 14 (Pineview Lane) did not demonstrate the same variability associated with runoff that the other sites demonstrated. Additionally, field visits to the site found the stream channel completely frozen. We believe monitoring during this period represents a pool of water below the ice during the winter.

Summer concentrations occur at River Miles 0.6 through 7.3 and River Mile 11.4. River mile 10.3 sits downstream of a wetland complex. Water stored and subsequently discharged from the wetland may be diluting concentrations at this site during base flow.

Four day average concentration time series for each for the six logged sites are presented in Appendix C.
6.2.2.2 Acute Exceedances

Figure 6.7 presents daily maximum concentrations at the six logged sites. Only two sites demonstrated acute exceedances including Zane (RM 7.3) and the outlet (RM 0.3). Zane Avenue had long durations above the acute standard in the winter, lasting thorough mid-March. Acute violations did not occur after spring rains arrived and snow pack was lost from the watershed. Four day average concentration time series for each for the six logged sites are presented in Appendix C.
6.3 GROUND WATER QUALITY

Ground water contributions to surface waters can constitute a significant portion of surface water loads for dissolved substances such as total dissolved solids or chloride. However, groundwater interactions with surface waters in the Shingle Creek watershed have not been thoroughly studied. The USGS completed a water quality assessment of groundwater quality in the Shingle Creek watershed and surrounding areas in 1996 (Andrews et al. 1996). Thirty shallow groundwater wells were installed, sampled and analyzed for 240 compounds including chloride. Chloride concentrations ranged from 4.3 to greater than 370 mg/L. Prior samples taken residential areas of the Anoka Sand Plain reported a substantially less median concentration of 26 mg/L (Anderson 1993). The spatial distribution of chloride concentrations in groundwater in the Shingle Creek watershed is presented in Figure 6.8.

Figure 6.7. Daily Maximum Chloride Concentrations Based On The Conductivity Chloride Relationships.
Figure 6.8. Chloride Concentrations in Groundwater Wells in the Shingle Creek Watershed and Surrounding Areas. Figure was adapted from Andrews 1996.

To assess loads to source waters, base flows were determined using the flow record. Once base flows were determined, concentrations were selected from each monitoring site during those flow periods after a long dry period. Incremental inflows and associated concentrations are presented in Table 6.4. Stream concentrations chosen were from grab samples collected on August 8, 2003.
Table 6.4. Incremental Inflow and Associated Concentrations and Daily Loads

<table>
<thead>
<tr>
<th>Site</th>
<th>Incremental Inflow (cfs)</th>
<th>Stream Concentration (mg/L)</th>
<th>Inflow Concentration (mg/L)</th>
<th>Inflow Load (tons/day)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Pineview</td>
<td>0.5</td>
<td>42</td>
<td>42</td>
<td>0.06</td>
</tr>
<tr>
<td>SCSS1</td>
<td>0.5</td>
<td>140</td>
<td>238</td>
<td>0.32</td>
</tr>
<tr>
<td>SC04</td>
<td>1</td>
<td>100</td>
<td>80</td>
<td>0.22</td>
</tr>
<tr>
<td>SC03</td>
<td>1</td>
<td>190</td>
<td>280</td>
<td>0.75</td>
</tr>
<tr>
<td>SCI94</td>
<td>2</td>
<td>180</td>
<td>175</td>
<td>0.94</td>
</tr>
<tr>
<td>SC00</td>
<td>0.7</td>
<td>200</td>
<td>257</td>
<td>0.48</td>
</tr>
<tr>
<td>Total</td>
<td>5.7</td>
<td>--</td>
<td>--</td>
<td>2.8</td>
</tr>
</tbody>
</table>

Some portion of the groundwater chloride is likely the result of natural sources including rock mineralization. Background conditions are difficult to identify but several studies may shed some light on the issue. The USGS sampled 992 wells in the Upper Mississippi River watershed where chloride concentrations ranged from 1-50 mg/L (Andrews et al, 1996). Chloride concentrations measured in groundwater wells in residential areas of the Anoka Sand Plain had a median concentration of 26 mg/L. Concentrations in ground water around Shingle Creek were higher than reported values in either of these two studies.
7.0 Linking Water Quality Targets and Sources

7.1 INTRODUCTION

A key aspect of a TMDL is the development of an analytical link between loading sources and receiving water quality. This analysis involves the solution of the equation for loading capacity as a function of wasteload allocation (WLA), load allocation (LA), margin of safety (MOS), and seasonal variation (SV).

\[
\text{TMDL} = \sum \text{WLA} + \sum \text{WLA} + \text{MOS}
\]

7.2 SELECTION OF MODELS AND TOOLS

An empirical approach was used to develop the chloride TMDL for Shingle Creek. The first step in the load allocation was using the analytical data collected in the watershed to identify flow conditions and seasons where the greatest occurrence of exceedances occurred. Target and measured loads were used to empirically develop load and wasteload allocations needed to meet water quality standards for chloride in Shingle Creek.

7.3 STREAM LOADS

7.3.1 Monitoring Year (2002-2003)

To assess stream loads, daily flow and load duration curves were developed for each of the sites with conductivity and flow data from December 1, 2002 to August 31, 2003. Flow duration curves are used to describe the frequency and occurrence of specific flow rates over a period of time. For example, a discharge of 5 cfs at an 80% flow interval tells us that the stream had a flow rate of 5 cfs or greater, 80% of the time. This results in breaking down the flow intervals from flood conditions (<1% interval) to dry conditions (90% interval). The real advantage to this approach is that data is presented across all the flow regimes and not restricted to a design flow...
criteria. This is essential since nonpoint source pollution is driven by runoff events and needs to be evaluated across all flow regimes.

Figure 7.1 presents the flow duration curve for the outlet of the watershed (RM 0.3). Flows ranged from approximately 2 cfs to over 600 cfs. All flow duration curves are presented in Appendix D.

![Flow Duration Curve](image)

**Figure 7.1. Flow Duration Curve for the Outlet of the Watershed (RM 0.3).**

These data are then used to develop a load duration curve for chloride (Figure 7.2). Flow intervals are described on the figure as ranging from dry to very high runoff conditions. Load violations occurred over the entire flow regimes at the outlet except at very high flows. Load duration plots for all sites can be found in Appendix D.
Load durations can be plotted seasonally to better understand violations on a seasonal basis across flow regimes. Seasonal load duration plots for all sites can be found in Appendix D.

Winter (December 1 through March 31) load violations (December 1 through March 31) occurred across all of the flow regimes (Figure 7.3).

Figure 7.2. Load Durations for the Shingle Creek Outlet (RM 0.3).

Figure 7.3. Winter (December 1 through March 31) Load Durations for the Shingle Creek Outlet (RM 0.3).
Spring (April and May) load violations occurred during the low flows (Figure 7.4). High flows offered enough dilution capacity or were late enough that the salt sources were depleted.

![Spring Chronic Load Duration](image1)

Figure 7.4. Spring (April and May) Load Durations for the Shingle Creek Outlet (RM 0.3).

Summer (June 1 through August 31) load violations did not occur (Figure 7.5). However, very dry periods had loads approaching the standard suggesting that ground water is close to the standard concentration of 230 mg/L.

![Summer Chronic Load Duration](image2)

Figure 7.5. Summer (June 1 through August 31) Load Durations for the Shingle Creek Outlet (RM 0.3).
Seasonal violation occurrences across the flow regimes are summarized in Table 7.1.

<table>
<thead>
<tr>
<th>Site</th>
<th>Winter</th>
<th>Spring</th>
<th>Summer</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Low</td>
<td>Medium</td>
<td>High</td>
</tr>
<tr>
<td>SC00</td>
<td>Yes</td>
<td>Yes</td>
<td>Yes</td>
</tr>
<tr>
<td>SCI94</td>
<td>Yes</td>
<td>Yes</td>
<td>Yes</td>
</tr>
<tr>
<td>SC03</td>
<td>Yes</td>
<td>Yes</td>
<td>Yes</td>
</tr>
<tr>
<td>SC04</td>
<td>Yes</td>
<td>Yes</td>
<td>Yes</td>
</tr>
<tr>
<td>SCSS1</td>
<td>--</td>
<td>Yes</td>
<td>No</td>
</tr>
<tr>
<td>SCPine</td>
<td>Yes</td>
<td>Yes</td>
<td>Yes</td>
</tr>
</tbody>
</table>

### 7.3.2 USGS Data

Additionally, we analyzed data collected by the USGS at the Queen Avenue Bridge from May of 1996 to December of 1998. The winter of 1996-1997 was a heavy snow year with 72.1 inches of snowfall. Exceedances still occurred across the entire winter except for the extremely high flows which probably represent late spring snowmelt (Figure 7.6).

![Winter Chronic Load Duration Quee Avenue 1996-97](image)

**Figure 7.6. Winter (December 1996 through March 31, 1997) Load Durations for Shingle Creek at the Queen Avenue Bridge.**

The winter of 1997-1998 was slightly below the average snowfall of 56 inches at 45 inches. Once again, the same pattern emerges where exceedances occur over the entire monitoring period (Figure 7.7). During this winter sampling period, high flows also demonstrated exceedances.
7.3.3 Reductions

Another way to analyze the data includes assessing the reductions needed for each daily load to reach the standard. The reductions needed to meet the standard during the monitoring year of 2002-2003 had a maximum of 72% and occurred during high flow periods (Figure 7.7). All flow categories had loads that required a reduction greater than 60%.

Figure 7.6. Winter (December 1997 through March 31, 1998) Load Durations for Shingle Creek at the Queen Avenue Bridge.

Figure 7.7. Percent Reductions Identified to Bring Individual Loads Below the Standard.
For comparison purposes, we also analyzed data collected at the Queen Avenue station by the USGS during the 1996-1997 and 1997-1998 winters. The winter of 1996-1997 was a heavy snow year with 72.1 inches of snowfall. Necessary reductions were as high as 59% (Figure 7.8). The winter of 1997-1998 was slightly below the average snowfall of 56 inches at 45 inches and required reductions as high as 62% with the greatest needed reductions in the 40% to 100% flow categories (Figure 7.9).

Figure 7.8. Percent Reductions Identified to Bring Individual Loads Below the Standard.

Figure 7.9. Percent Reductions Identified to Bring Individual Loads Below the Standard.
In our case, the monitored year turned out to be a worst-case year in that the amount of salt used compared to the precipitation was high resulting in a lowered dilution capacity because less water was on the watershed in the form of snow pack. This is demonstrated by the greatest load reductions needed in the lightest snow year. The largest snow year required the smallest percent reductions.
8.0  TMDL Allocation

8.1  TMDL

Critical conditions defined for the load and wasteload allocations were defined as all winter flow conditions. However, because the chloride loading functions as a non-point source issue in the Shingle Creek watershed, it is inappropriate to define the TMDL as a single number since the TMDL as developed is entirely dependant on the daily flow and concentration, which is highly dynamic. To this effect, the TMDL is represented by an allowable daily load across all flow regimes as is demonstrated in Figure 8.1. To determine acceptable loads under the critical flow regimes, chronic standard concentrations were multiplied by the flow at each interval.

![Winter Load Duration Chart](image)

Figure 8.1. Total Maximum Daily Load Across Flow Exceedances for Shingle Creek. Data used to calculate the load duration curve was from December 1996 thorough March 2003.

To better facilitate implementation, TMDL guidance suggests that alternate expressions of the TMDL can be applied where appropriate. In this case, the TMDL is represented as a percent reduction across the flow regimes needed to meet the standard (Table 8.1). The TMDL is set
such that all of the loads would come into compliance. In other words, the reduction is set to the highest required reduction based on the monitoring data.

**Table 8.1. TMDL for Chlorides in Shingle Creek as Represented by a Percent Reduction.**

<table>
<thead>
<tr>
<th>Critical Condition</th>
<th>Wasteload Allocation (percent reduction)</th>
<th>Load Allocation (percent reduction)</th>
<th>Margin of Safety (percent reduction)</th>
<th>TMDL (percent reduction)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Winter Low Flow (60 to 100%)</td>
<td>60%</td>
<td>3%</td>
<td>Implicit</td>
<td>63%</td>
</tr>
<tr>
<td>Winter Runoff (60% to 0%)</td>
<td>67%</td>
<td>4%</td>
<td>Implicit</td>
<td>71%</td>
</tr>
</tbody>
</table>

1Assumed groundwater reductions with reductions of surface application of chloride (37% and 52% respectively). Total load reduction was based on an assumed stream load share of 8%. For example, a 37% load reduction on 8% of the load results in a 3% reduction of the entire load.

The TMDL can also be expressed as a set of daily equations derived from the load duration curve. Table 8.2 represents the TMDL for the 5th, 25th, 50th, 75th, and 95th flow duration intervals.

**Table 8.2. TMDL for Chlorides in Shingle Creek as Represented by Daily Loads.**

<table>
<thead>
<tr>
<th>Load Duration Interval</th>
<th>Wasteload Allocation (tons/day)</th>
<th>Load Allocation (tons/day)</th>
<th>Margin of Safety (tons/day)</th>
<th>TMDL (tons/day)</th>
</tr>
</thead>
<tbody>
<tr>
<td>5%</td>
<td>23.2</td>
<td>1.6</td>
<td>Implicit</td>
<td>24.8</td>
</tr>
<tr>
<td>25%</td>
<td>7.2</td>
<td>1.6</td>
<td>Implicit</td>
<td>8.8</td>
</tr>
<tr>
<td>50%</td>
<td>2.9</td>
<td>1.6</td>
<td>Implicit</td>
<td>4.5</td>
</tr>
<tr>
<td>75%</td>
<td>1.8</td>
<td>1.6</td>
<td>Implicit</td>
<td>3.4</td>
</tr>
<tr>
<td>95%</td>
<td>0.3</td>
<td>1.6</td>
<td>Implicit</td>
<td>1.9</td>
</tr>
</tbody>
</table>

1Assumed groundwater reductions with reductions of surface application of chloride (45% reduction).

### 8.2 LOAD ALLOCATION (LA) AND WASTELOAD ALLOCATION (WLA)

Because stormwater discharges are regulated under NPDES Phase II, allocations of chloride reductions are considered wasteloads and must be divided among permit holders. Although the cities hold individual permits, they are combined here to reflect their participation in the SCWMC.

To support determination of source load reductions needed to meet the standard, a thorough inventory of chloride sources was conducted. Table 8.3 outlines the sources and their overall contribution to chloride in the watershed.
Using the information provided, a stakeholder process was used to determine load allocations among users in the watershed. The stakeholders in the watershed agreed to work collectively to achieve a 71% reduction in chloride use to achieve the standard understanding that each stakeholder was working under unique financial, public safety and perception, and feasibility limitations. However, each stakeholder agreed to implement BMPs to the maximum extent practicable. This collective approach allows for greater reductions for some agencies and less for those with greater constraints. The collective approach is to be outlined in an implementation plan developed by the Shingle Creek Watershed Management Commission.

8.3 RATIONALE FOR LOAD AND WASTELOAD ALLOCATIONS

8.3.1 Rationale for Load and Wasteload Allocations

The allocations are based on evaluation of chloride and flow monitoring in Shingle Creek during 2002 and 2003. Monitoring, using conductivity as a surrogate measure of chloride, provided daily loads of chloride in the Shingle Creek watershed. Measured daily loads were then compared to acceptable loads across the suite of flows that occur in Shingle Creek providing the basis for the load allocations.

To determine acceptable loads under the critical flow regimes, the chronic standard concentration was multiplied by the flow at each interval. Measured loads can then be compared to standard loads to determine the percent difference between the values and ultimately the percent reduction needed to meet the standard. To develop the load allocations, critical flow period were identified on the flow duration curve, which included to 10% to 60% duration interval and the 60% to 90% duration interval. Load reductions are presented on Figure 8.2.

---

Table 8.3. Chloride Sources in the Shingle Creek Watershed.

<table>
<thead>
<tr>
<th>Assumed Sources</th>
<th>Total Chloride (tons)</th>
<th>Daily Load (tons/day)</th>
<th>Percent of Total</th>
</tr>
</thead>
<tbody>
<tr>
<td>Road Salt Cities</td>
<td>2,790</td>
<td>23.1</td>
<td>43%</td>
</tr>
<tr>
<td>Road Salt Hennepin County</td>
<td>1,660</td>
<td>13.7</td>
<td>26%</td>
</tr>
<tr>
<td>Road Salt MnDOT</td>
<td>858</td>
<td>7.1</td>
<td>13%</td>
</tr>
<tr>
<td>Road Salt Storage Facilities</td>
<td>290</td>
<td>2.4</td>
<td>5%</td>
</tr>
<tr>
<td>Private Application</td>
<td>463</td>
<td>3.8</td>
<td>7%</td>
</tr>
<tr>
<td>Residential</td>
<td>53</td>
<td>0.4</td>
<td>1%</td>
</tr>
<tr>
<td>Groundwater</td>
<td>335</td>
<td>2.8</td>
<td>5%</td>
</tr>
<tr>
<td>TOTAL</td>
<td>6,449</td>
<td>50.5</td>
<td>100%</td>
</tr>
</tbody>
</table>

Reduction based on groundwater returning to natural background levels of <50 mg/L.
The load allocation represents the groundwater portion of the stream chloride load. To determine groundwater load reductions, we assumed groundwater chloride was reduced linearly with surface reductions to a minimum of 50 mg/L, which is the assumed background chloride concentration. For example, a 51% reduction in chloride sources to groundwater would reduce the groundwater source by 37% since the reduction is only applied to the assumed non-background chloride load. The total load reduction was based on an assumed stream load share of 8%. A 37% load reduction on 8% of the entire load results in a 3% reduction of the entire load. It is also important to note that this reduction is considered a long-term effect since groundwater flushing will take many years to purge prior chloride additions.

![Winter Chronic Load Duration](image)

**Figure 8.2 TMDL Applied to the 2002-2003 Monitoring Season.** The red line represents the TMDL. The black line represents the loads across flow durations where the allocated load reductions would result in all of the measured loads meeting the standard.

### 8.3.2 Margin of Safety

The Margin of Safety - MOS - is implicit. The TMDL calls for a 71% reduction of chloride during all conditions. Much of the runoff results from the melting of roadside snow from previous snowfall events and therefore previous road salt applications. The 71% reduction was determined based upon the highest single exceedance of the WQS. This 71% is not the direct
result of a 71% excessive application of chloride, rather, it represents the cumulative impact of multiple events. However, since the cumulative impacts cannot be quantified at this time, MPCA believes using the 71% target is a conservative assumption that overestimates the chloride reduction needed to achieve WQSs.

As the overall 71% reduction is achieved, the salt burden held in the accumulated roadside snow from previous snows will be significantly reduced over the conditions that existed during the TMDL development winters. This compounding reduction (71% during all conditions) should ensure achieving water quality standards during future critical conditions (winter snowmelt and runoff).

8.4 SEASONAL AND ANNUAL VARIATION

8.4.1 Seasonal Variation

Conductivity and chloride data analyzed for this TMDL were collected from December 2002 through August 31, 2003. Data were analyzed seasonally including winter (December 1 through March 31), Spring (April 1 through May 31) and summer (June 1 through August 31). These periods reflect differences in the mass of chloride available since road salt is applied only during the snow and ice season. Fall will act much the same as summer since no application of chloride (road salt) occurs and the chloride source is groundwater. Winter and spring were evaluated separately since runoff is produced through different processes during these seasons. Winter runoff is primarily snowmelt resulting from warm periods and high sun intensity. Spring is primarily precipitation events. Since snow accumulates in snow piles adjacent to the roads, snowmelt can deliver runoff extremely high in chloride concentrations. These differences have been accounted for in the identification of the critical periods and allocations for each of the critical periods.

8.4.2 Annual Variation

Load allocations for this TMDL are based on monitoring from December 2002 through August 31, 2003. The better understand annual variability, load durations based on the chloride standard of 230 mg/L were compared for winter months for both the long-term record and analysis year
Figure 8.3. The two curves are almost identical. There is a difference in the 80 to 100% flow duration categories with the analysis year allowable load lower than the long-term allowable load. This is most likely due to utilizing data from a light snow/precipitation year where low flows were lower than normal. This could also be caused by an extended dry summer/fall period where groundwater contributions are less during the following winter.

To illustrate that the proposed reductions are protective of the standard in all years, we analyzed data collected by the USGS at the Queen Avenue Bridge from May of 1996 to December of 1998. The winter of 1996-1997 was a heavy snow year with 72.1 inches of snowfall. The winter of 1997-1998 was slightly below the average snowfall of 56 inches at 45 inches. These two years required a maximum reduction of 59% and 62% respectively (Figures 7.8 and 7.9). Based on this analysis the current TMDL would be protective of the standard in more average snow years. Additionally, TMDLs are often set to the most sensitive conditions or the “critical conditions”. In our case, the monitored year turned out to be a critical condition in that the amount of salt used compared to the precipitation was high resulting in a lowered dilution capacity because less water was on the watershed in the form of snow pack. Consequently, the TMDL appears to be protective of the critical conditions of the watershed.
8.5 FUTURE GROWTH

Most of the currently undeveloped or lightly developed areas of northern Brooklyn Park, southeastern Maple Grove, and northwestern Plymouth are expected to be developed by 2020. Growth is expected to include residential, commercial, and industrial development. Invariably, some of this development will include roads and ultimately increased amounts of chloride based deicer use in the watershed. Areas of northern Brooklyn Park that will be developed are mostly outside of the watershed and drain directly to the Mississippi River. Increases in development are expected to be relatively small since the watershed is essentially fully developed. Expected development in Maple Grove would impact Shingle Creek directly while expected development in Plymouth would impact Bass Creek.

Since the changes are relatively small and the majority of roads associated with this development would be low speed, residential roads, only small increases in chloride use would be expected. Any policies or BMPs prescribed by this TMDL would be implemented on the new roads and developed areas. Consequently, provisions for new growth is built into the TMDL as a part of the adaptive management approach.
9.0 Public Participation

9.1 INTRODUCTION

As a part of the strategy to achieve implementation of the necessary allocations, the SCWMC sought stakeholder and public engagement and participation regarding their concerns, hopes, and questions regarding the development of the TMDL. Specifically, meetings were held for a Technical Advisory Committee representing key stakeholders and local experts. Additionally, the SCWMC held a series of stakeholder meetings focused on implementation of the TMDL requirements.

The SCWMC maintains an interactive website. The TMDL and all related material were posted on this website. Stakeholder and other public meeting notices were posted on this website. The NBC News affiliate, KARE 11, did a news piece on road salt (chloride) featuring Shingle Creek. This news piece reached an audience of approximately 1.5 million households. The news piece is/was posted on SCWMC’s website.

9.2 TECHNICAL ADVISORY COMMITTEE

A technical advisory committee was established so that interested stakeholders could be involved in key decisions in developing the TMDL. Stakeholders represented on the Technical Advisory Committee include the 10 local cities, Hennepin County, Mn/DOT, Minnesota DNR, the Metropolitan Council, the USGS and the Minnesota Pollution Control Agency. All meetings were open to interested individuals and organizations. Technical Advisory committee meetings were held at regular intervals during the development of the TMDL.
9.3 STAKEHOLDER MEETINGS

A detailed stakeholder process was conducted for the Shingle Creek Chloride TMDL that included meetings and work sessions to identify activities (BMPs) that may be implemented to address chloride exceedances in Shingle Creek. The stakeholder process focused on the agencies responsible for winter road maintenance and included member cities of the SCWMC, Mn/DOT, and Hennepin County. The stakeholder process focused on these groups because of the inherent need to address both public safety and the environmental concerns of deicing activities. The necessary reductions in chloride will be implemented primarily by these agencies and will ultimately change the way roads are maintained for winter snow and ice conditions. Additionally, a vast amount on knowledge resides in this group concerning the newest technologies, the feasibility of implementing BMPs, and the extent of service required to protect public safety. Stakeholder meetings were held on the following dates:

- February 4, 2005
- February 25, 2005
- April 1, 2005
- May 6, 2005

9.4 PUBLIC MEETINGS

The SCWMC maintains an interactive website. The TMDL and all related material were posted on this website. Stakeholder and other public meeting notices were posted on this website. The NBC News affiliate, KARE 11, did a news piece on road salt (chloride) featuring Shingle Creek. This news piece reached an audience of approximately 1.5 million households. The news piece is/was posted on SCWMC’s website.

The TMDL was noticed on the State of Minnesota’s register with a 30-day public comment period.
10.0 Implementation

10.1 DEVELOPMENT OF THE IMPLEMENTATION PLAN

The activities and BMPs identified in the implementation plan are the result of a series of stakeholder working-meetings led by the Shingle Creek Watershed Management Commission. The meetings focused on the discussion of the TMDL requirements, BMPs and technologies available to address chloride, public safety, and the feasibility of implementing the activity.

Additionally, MnDOT developed a “Best Available Technologies” report outlining the state of BMPs in six categories. That report is attached as appendix H. The MnDOT report and the stakeholder discussions during the load reduction/implementation development, identified BMPs ranked the smallest level of implementation to the greatest level of implementation. The ranking was as follows:

No BMP<Minimum BMP<Maximum Extent Practicable<Best Available Technology

The load allocations in this TMDL represent aggressive goals for chloride reductions with the added challenge of addressing public safety and expectation. Consequently, implementation will be conducted using adaptive management principles. Adaptive management is appropriate because it is difficult to predict the chloride reduction that will occur from implementing strategies with the paucity of information available to demonstrate expected reductions. Continued monitoring and “course corrections” responding to monitoring results are the most appropriate strategy for attaining the water quality goals established in this TMDL while maintaining required levels of public safety.
10.2 IMPLEMENTATION FRAMEWORK

Member cities of the SCWMC, Mn/DOT, and Hennepin County have all agreed to identify and implement BMPs focused on reducing chloride use in the Shingle Creek watershed. Stakeholder meetings focused on the Cities’ current activities and identification of activities that can be added to address the needed load reductions in the Chloride TMDL. The topics for the meeting included:

1. Product Application Equipment and Decisions
2. Product Stockpiles
3. Product Type and Quality
4. Operator Training
5. Clean-up and Snow Stockpiling
6. Ongoing Research into Salt Alternatives

During the stakeholder process, each of the cities discussed their current methodologies and practices for winter road maintenance and identified those areas where improvements could be achieved in each of the six identified categories. Results of these discussions are included in Table H1 through H6 in Appendix I. The following section is a general summary of the activities to be implemented under each of the six categories.

10.3 IDENTIFIED REDUCTION STRATEGIES

The SCWMC will work through the above framework to encourage implementation of the following strategies. Although the SCWMC will be the lead on the implementation of the Chloride TMDL, individual stakeholders will be ultimately responsible for implementing the identified BMPs. These activities will be tracked by the MPCA as part of the NPDES Phase II Permits that all of the stakeholders hold. The NPDES Phase II permits are BMP based calling for BMPs at the Maximum Extent Practicable (MEP) level to achieve applicable water quality standards. Mn/DOT’s reduction strategies are covered in the BAT Report included in Appendix H.
10.3.1 Product Application Equipment and Decisions

In many cases, less road salt can be used without compromising public safety. To avoid over application, standards can be established for application rates that account for pavement temperature ranges and timing. Newer technologies such as pre-wetting and anti-icing can result in the same results while using significantly less product. Pre-wetting of salt refers to applying water, or some other liquid agent such as magnesium chloride, to the salt either prior to or during application of the material. Pre-wetting reduces the amount of scatter and loss of material, ultimately reducing the usage amounts. To this end, the stakeholders in the watershed have agreed to incorporate the following practices:

1. Annually calibrate spreaders
2. Use the Road Weather Information Service (RWIS) and other sensors such as truck mounted or hand held sensors to improve application decisions such as the amount and timing of application
3. Evaluate new technologies such as prewetting and anti-icing as equipment needs to be replaced. These technologies will be adopted where feasible and practical.
4. Investigate and adopt new products (such as Clear Lane, a commercially available pretreated salt) where feasible and cost effective

The estimated cost of implementing this activity will vary based on the technologies. Some examples include:

- Dry tailgate spreader: $3,000
- Prewetting: $6,000
- Spreader: $9,000
- Epoke spreaders: $60,000
- Brine storage system: $25,000
- Salt: $34/ton; Clear Lane: $39/ton + $5/ton delivery

10.3.2 Deicer Stockpiles

Another source of chloride is runoff from salt storage facilities. The stakeholders agreed to cover all product stockpiles and store them on impervious surfaces. Additionally, stakeholders will maintain general good-housekeeping policies associated with the handling of road salt to
minimize the potential for wash-off of excess or spilled salt. There is no additional cost expected for this activity.

10.3.3 Operator Training

Stakeholders identified operator training as a primary area that could result in significant reductions in road salt use. One aspect of the training is to discuss the environmental concerns with the public safety issues to reinforce the concept of using the least amount of product necessary to maintain public safety. The stakeholders agreed to have annual training that may include outside support such as LTAP (Local Technical Assistance Program) or vendor training on the appropriate use of technologies or products. The estimated cost of this activity is $1,000 for staff time annually per LGU.

10.3.4 Cleanup and Snow Stockpiling

Snow disposal can be a concern, especially in areas where snow cannot be pushed off the side of the road. Snow plowed directly streamside can leak high concentrations of chloride into the stream. This is of special concern during base flow resulting in increased chloride concentrations. Although little snow hauling occurs in the Shingle Creek watershed, the stakeholders agreed to stockpile snow away from sensitive areas. A sensitive area is defined as directly streamside, on slopes greater than 6%, or near a wetland or storm sewer inlet. All stakeholders also agreed to sweep City streets as soon as possible in late winter to remove as much residual product as possible. There is no additional cost expected for this activity.

10.3.5 Ongoing Research into Salt Alternatives

Technologies associated with winter road maintenance are constantly changing based on the needs of the industry. Due to the changing technologies, there is a need to keep informed on new practices, technologies, and products that can ultimately protect public safety and the environment. All of the stakeholders will evaluate the technologies on an annual basis and implement the most appropriate technologies where feasible. The estimated cost of this activity is $2,000 for staff time annually per LGU, plus the cost of any technologies implemented.
10.3.6 SCWMC Activities

The SCWMC has agreed to take the lead on public education and private applicator education. The following activities will be conducted by the SCWMC.

Coordinate an Annual Commercial Applicator Workshop

The purpose of the workshop is to discuss salt usage, application techniques, and storage issues, product type and alternatives, and other technologies so that commercial applicators are informed. The estimated cost of this activity is $1,000 annually.

Private Applicator Education

Education of private applicators (commercial, industrial, and residential) and homeowners can help reduce chloride based deicer use in the watershed. Some educational materials have been developed by Canadian agencies regarding private use of chloride-based deicers. Private applicator education will include development of brochures, newsletters, website pieces, and presentations to educate private applicators on chloride issues in the watershed. The estimated cost of this activity is $1,500 annually.

Permit Requirements

The commission will incorporate private (commercial) snow management rules for reducing chloride use and include chloride reduction in the Commission’s project review program. One requirement may be the development of a salt management plan for individual commercial properties. The commission will develop a template for the salt management plan. The estimated cost of this activity is $2,000.

Conduct Official Education

There is a need for City, County, and State officials to understand the TMDL and the proposed implementation activities so that they can effectively balance the public safety issues with the
environmental risks. The SCWMC will inform the appropriate officials with the necessary information. The estimated cost of this activity is $1,000 annually.

Monitoring

Monitoring of chloride and conductivity at two locations is already incorporated into the Commission’s annual monitoring activities. The estimated cost of this activity is $3,000 annually.

Public Education and Outreach

One measure that may allow for reductions in usage of deicing chemicals is to increase public knowledge of the environmental effects of road salt and ultimately gain public acceptance in lowering driving speeds during icy conditions. Another effect education can have is lowering public expectations for snow removal and deicing. This task will educate the public to help manage expectations and identify the need for chloride reductions. Activities may include newsletter articles, brochures, website pieces and presentations. The estimated cost of this activity is $3,000 annually.

Annual Report on Monitoring and Activities

An annual report on salt reduction activities is necessary under the adaptive management guidelines established in the TMDL. This report will provide the Cities’ with necessary information for their annual NPDES reports. The report will track BMP scheduling, implementation, O & M and environmental condition monitoring data to evaluate activity effectiveness. The estimated annual cost of this activity is $5,000.

City Salt Management Plans

The implementation plan asks the Cities to develop and maintain a City Salt Management Plan. Many Cities already have these, but a template is needed to easily compare activities between Cities. A template will reduce the Cities’ workload and provide an easily amendable plan for
reducing salt use. The SCWMC will develop a template for the City Salt Management Plans at an estimated cost of $3,000.

**10.3.7 Monitoring Implementation of Policies and BMPs**

The SCWMC will evaluate progress toward meeting the goals and policies outlined in the Second Generation Plan in their Annual Report. Success will be measured by completion of policies and strategies, or progress toward completion of policies and strategies. The Annual Report will be presented to the public at the Commission’s annual public meeting. The findings of the Annual Report and the comments received from the member cities and the public will be used to formulate the work plan, budget, Capital Improvement Program (CIP) and specific measurable goals and objectives for the coming year as well as to propose modifications or additions to the management goals, policies, and strategies.

**10.3.8 Follow-up Monitoring**

The SCWMC monitors water quality at two stations in the watershed (Zane Ave. and Humboldt Ave. near the outlet). Upon the initiation of this TMDL study, the SCWMC has increased monitoring at these two stations to include grab samples of chloride and collection of conductivity at 15-minute intervals. These data will be used to track effectiveness of BMP implementation. Results will be included in the Commission’s annual water quality monitoring report.

The SCWMC has agreed to take the lead on monitoring and tracking the effectiveness of activities implemented to reduce chloride in Shingle Creek.
11.0 Reasonable Assurance

11.1 INTRODUCTION

When establishing a TMDL, reasonable assurances must be provided demonstrating the ability to reach and maintain water quality endpoints. Several factors control reasonable assurances including a thorough knowledge of the ability to implement BMPs as well as the overall effectiveness of the BMPs. This TMDL is unique in that it requires maintaining a balance between protecting the beneficial use of the water body and public safety. Additionally, the scientific understanding of BMP effectiveness for chloride is still young and research must account for changes in public safety. To address these issues, adaptive management will be implemented to protect water quality without sacrificing public safety. As research and understanding on the potential BMPs begin to solidify our understanding of their ability to maintain public safety and protect the beneficial uses of the water body, actions and management plans will be changed to incorporate these advances. However, there are some BMPs and policies that can be addressed now to improve water quality conditions in Shingle Creek. Additionally, there is a need to begin implementation and monitor the effectiveness of these BMPs in meeting the load allocations.

11.2 THE SHINGLE CREEK WATERSHED MANAGEMENT COMMISSION

The Shingle Creek Watershed Management Commission was formed in 1984 using Joint Powers Agreements developed under authority conferred to the member communities by Minnesota Statutes 471.59 and 103B.201 through 103B.251. The Commissions’ purpose is to preserve and use natural water storage and retention in the Shingle Creek watershed to meet Surface Water Management Act goals. 

The Shingle Creek and West Mississippi Watershed Management Commissions – briefly explain that these two have cooperated to plan and act.
The Metropolitan Surface Water Management Act (Chapter 509, Laws of 1982, Minnesota Statute Section 473.875 to 473.883 as amended) establishes requirements for preparing watershed management plans within the Twin Cities Metropolitan Area. The law requires the plan to focus on preserving and using natural water storage and retention systems to:

- Improve water quality.
- Prevent flooding and erosion from surface flows.
- Promote groundwater recharge.
- Protect and enhance fish and wildlife habitat and water recreation facilities.
- Reduce, to the greatest practical extent, the public capital expenditures necessary to control excessive volumes and rate of runoff and to improve water quality.
- Secure other benefits associated with proper management of surface water.

Minnesota Rules Chapter 8410 requires watershed management plans to address eight management areas and to include specific goals and policies for each. Strategies and policies for each goal were developed to serve as a management framework. To implement these goals, policies, and strategies, the Commission has developed the Capital Improvement Program and Work Plan discussed in detail in the Second Generation Plan (SCWMC 2004).

The philosophy of the Joint Powers Agreement is that the management plan establishes certain common goals and standards for water resources management in the watersheds, agreed to by the ten cities having land in the watersheds, and implemented by those cities by activities at both the Commission and local levels. TMDLs developed for water bodies in the watershed will be used as guiding documents for developing appropriate goals, policies, and strategies and ultimately sections of the Capital Improvement Program and Work Plan.

The SCWMC is committed to improving water quality in the Shingle Creek watershed. To this end, the SCWMC has recently completed a water quality management plan. The Shingle Creek and West Mississippi Watershed Management Commissions’ Water Quality Plan (WQP) is intended to help achieve a Second Generation Management Plan goal of protecting and improving water quality. A number of activities are proposed in the Management Plan over the
next ten years, including developing individual management plans for major water resources. One specific activity identified in the plan was the completion and implementation of the chloride TMDLs for Shingle Creek.

The Shingle Creek Water Quality Plan (WQP) is intended to:

- Set forth the Commissions’ water quality goals, standards, and methodologies in more detail than the general goals and policies established in the Second Generation Management Plan.
- Provide philosophical guidance for completing water resource management plans and TMDLs; and
- Provide direction for the ongoing water quality monitoring programs that will be essential to determining if the TMDLs and implementation program are effectively improving water quality.

The Shingle Creek and West Mississippi Watershed Commissions’ Water Quality Implementation Plan is composed of four parts:

- A monitoring plan to track water quality changes over time;
- Detailed management plans for each resource to lay out a specific plan of action for meeting water quality goals;
- A capital improvement plan; and
- An education and public outreach plan.

This Implementation Plan charts the course the Commissions will take to meet their Second Generation Management Plan goals to protect and improve water quality and meet Commission and State water quality standards. While the Plan lays out a series of activities and projects, implementation will occur as the Commissions’ and cities’ budgets permit.

The Commissions have received significant grant funding from the Minnesota Pollution Control Agency, the Board of Water and Soil Resources, the Metropolitan Council, and the Department of Natural Resources to undertake planning and demonstration projects. The Commissions intend to continue to solicit funds and partnerships from these and other sources to supplement
the funds provided by the ten cities having land in the two watersheds. It is expected that the Commissions will continuously update their annual Capital Improvement Programs (CIPs) as a part of their annual budget process.

11.3 NPDES MS4 STORMWATER PERMITS

NPDES Phase II stormwater permits are in place for each of the member cities in the watershed as well as Hennepin County and Mn/DOT. Under the stormwater program, permit holders are required to develop and implement a Stormwater Pollution Prevention Program (SWPPP; MPCA, 2004). The SWPPP must cover six minimum control measures:

- Public education and outreach;
- Public participation/involvement;
- Illicit discharge, detection and elimination;
- Construction site runoff control;
- Post-construction site runoff control; and
- Pollution prevention/good housekeeping.

The permit holder must identify BMPs and measurable goals associated with each minimum control measure. The EPA requires that stormwater sources of a pollutant addressed in a TMDL must be treated as a wasteload allocation (i.e., a point source). Under the NPDES provisions, the permit will require addressing load allocations as either an effluent limit or as BMPs or other similar requirements. Stormwater loads in this TMDL are allocated among the permit holders while combining the cities. Combination of the cities maintains the watershed approach of the Shingle Creek Watershed Management Commission.
11.4 EFFICACY OF BEST MANAGEMENT PRACTICES

Source reduction strategies and BMPs are starting to be implemented in the Snow Belt region and have shown promise in reducing chloride loads while maintaining public safety. These practices and policies are adoptable by local resource managers and stakeholders.

**Improved Equipment**
Improved technologies have demonstrated reductions in road salt usage and ultimately reduced costs in acquiring material. Prewetting material has been linked to reductions of up to 30% (www.saltinstitute.org).

**Deicing Alternatives**
Numerous deicing alternatives exist, however the majority of these carry other water quality impacts. Most of the alternatives include chloride based deicers and are often more toxic than sodium chloride. Other, “organic” alternatives typically have a high BOD. Shingle Creek is currently impaired due to low oxygen. Consequently, deicers that increase BOD are not a feasible alternative at this time in the Shingle Creek watershed.

11.5 MONITORING

The SCWMC has agreed to take the lead on monitoring and tracking the effectiveness of activities implemented to reduce chloride in Shingle Creek. The monitoring effort is a key aspect of adaptive management in that an annual evaluation of chloride data from Shingle Creek provides for an assessment of BMP effectiveness. Evaluation of the monitoring data will be included in the Shingle Creek Annual Monitoring Report.
12.0 References


Minnesota Department of Natural Resources State Climatology Office, 2000.


MPCA 2004b. 2004 Guidance Manual for Assessing the Quality of Minnesota Surface Waters for the Determination of Impairment


